



Research Journal of
**Environmental
Sciences**

ISSN 1819-3412



Academic
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Adsorption Kinetics: Bioremoval of Trivalent Chromium from Tannery Effluent by *Aspergillus* sp. Biomass

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Abstract: Removal of trivalent chromium from tannery effluent by microbial biomass generated from fermentation industry was carried out. Adsorption studies were conducted in batch mode at different adsorbent dosage, pH, initial concentration and contact time. The 94.6 and 64.48% of Cr (III) removal from aqueous solution and tannery effluent was observed at pH 4 by 1% adsorbent dosage, respectively. A comparison of kinetic models and correlation coefficient ($R^2 = 0.998$) indicated that Ho and Mckay kinetic model correlates well with the experimental kinetic data. The surface properties of the adsorbent before and after chromium sorption were characterized by X-ray diffraction and infrared technique. It was concluded that Cr (III) is significantly removed by *Aspergillus* biomass and can be used economically for the removal of Cr (III) from aqueous solution and tannery effluent.

Key words: Trivalent chromium, microbial biomass, tannery effluent, *Aspergillus* biomass, adsorption kinetics

INTRODUCTION

Trivalent chromium is an important pollutant introduced into natural waters by a variety of industrial waste water including those from the textile, leather tanning, electroplating and metal finishing industries (Gaballah and Kilbertus, 1998). Tannery industry is one of the oldest and fastest growing industries in India and there are about 2,161 tanneries excluding cottage industries, which discharge annual of 9,420,000 m³ wastewater into the environment.

Tannery industries use a great number of chemical agents and produce an enormous volume of residual waters and solid residues (Mohan *et al.*, 2006). The effluent from the chromium stream consists primary of a saline solution of spent chromium and residual fungicide at a pH between 3 and 5 (Tobin and Roux, 1998). In the world, the most usual technology for the treatment of the liquid residues generated by the leather industries constitutes of physico-chemical conventional treatment methods such as ion exchange (Tiravanti *et al.*, 1997; Rengaraj *et al.*, 2003; Petruzzelli *et al.*, 1995), reduction (Seaman *et al.*, 1999), chemical precipitation (Zhou *et al.*, 1993). Membrane (Shalan *et al.*, 2001; Kozlowski and Walkowiak, 2002), electrochemical precipitation (Kongsricharoen and Polprasert, 1996), photocatalytic reduction (Testa *et al.*, 2004), adsorption (Srivastava *et al.*, 1997; Gupta *et al.*, 1999) and biosorption (Gupta *et al.*, 2001; Arvindhan *et al.*, 2004).

The precipitation, oxidation/reduction, lime neutralization have traditionally been the most commonly used. Although, these technologies are quite satisfactory in terms of purging chromium from water, they produce solid sludge containing toxic compounds. Therefore, the need exists for a simple and robust treatment strategy that addresses local resources and constraints. Biosorption may be an effective and versatile method for removing chromium, particularly when combined with appropriate

regeneration steps. This solves the problems of sludge disposal and renders the system more economically viable, especially because low cost adsorbents are abundant in nature (Ahluwalia and Goyal, 2007). In this study, microbial waste biomass of *Aspergillus* sp. (MB1) obtained from fermentation industry was used for removal of Cr (III) from aqueous solution and tannery effluent. Adsorption kinetics was studied at varying adsorbent dosage, pH and initial Cr (III) concentrations.

MATERIALS AND METHODS

The chromium removal study was carried out in the Department of Biotechnology and Environmental studies, Thapar University, Patiala (Punjab) from January to July 2008.

Collection and Characterization of Microbial Biomass and Tannery Effluent

Microbial waste biomass of *Aspergillus* sp. (MB1), which is generated as a byproduct of pharmaceutical fermentation industry for the production of certain antibiotics was collected from Ranbaxy (fermentation industry) Paonta Sahib, Himachal Pradesh, India. This biomass was characterized for physical and chemical parameter such as CHN analysis, pH, moisture, ash content, bulk density and calorific value as per the procedure (MacDonald *et al.*, 1996).

Tannery effluent was collected from A.V. Tanneries, Kapurthala, Punjab, India. Characterization of effluent was performed as per procedure described elsewhere (Clesceri *et al.*, 1998; APHA, 1998).

Surface Characterization of Microbial Biomass

X-Ray Diffraction (XRD) pattern and Fourier Transform Infra-Red (FTIR) spectroscopy studies were carried on microbial biomass to determine the adsorption mechanism. A structural change of biomass before and after reacting with Cr (III) was determined by using an X-ray diffractometer (Philips PW 1830). The FTIR spectrum of native and chromium treated biomass was obtained after drying the biomass at 70°C. The finely powdered samples were encapsulated with potassium bromide to prepare translucent sample disks and spectra were recorded by fourier transform infrared spectroscope (Hartmann and Braun, BOMEM Canada).

Adsorbent Preparation

Microbial biomass was oven dried at 80°C for two days and ground in a blender and sieved to pass through a 2.0 mm sieve in order to obtain uniform particle size (2 mm) and washed several times with double distilled water and dried before being used as a sorbent for removal of Cr (III) from aqueous solution and tannery effluent.

Preparation of Aqueous Chromium Solution

Aqueous solution of chromium was prepared by dissolving 0.38 g of Cr₂O₁₂S₃ in 100 mL distilled water to obtain a concentration of 1000 mg L⁻¹. Stock solution and tannery effluent were diluted with distilled water to obtained working solution in the concentration range of 5 to 50 mg L⁻¹ and pH 2 to 5 of the solution was adjusted with 1N NaOH and 1N HCl.

Batch Adsorption Studies

Batch experiments were conducted with microbial waste biomass (MB1) in Erlenmeyer flask (250 mL) using different biomass dosage (0.25-2% of biomass) and aqueous solution (100 mL) of Cr (III) (25 mg L⁻¹). The suspension was shaken at 120 rpm, 28°C for 24 h. Similarly, experiments were conducted at different pH (2 to 5) and chromium concentration (5-50 mg L⁻¹). The residual concentration of chromium in aqueous solution was measured by (AAS) atomic absorption

spectrophotometer (GBC 932AA, Australia) with an air-acetylene flame using Cr-hallow cathode lamp having sensitivity limit of $0.05 \mu\text{g mL}^{-1}$. Working standard solution of chromium was prepared from stock (1000 mg L^{-1}) procured from Acros Organic Ltd., New Jersey, USA. Triplicates of each sample were analyzed. The mean value and relative standard deviation as given by AAS were recorded.

Chromium uptake (q_e) was calculated as follows:

$$q_e = \frac{(C_i - C_f)V}{m} \quad (1)$$

where, C_i and C_f are the initial and final Cr (III) concentrations in mg L^{-1} , respectively, V is the volume of the Cr (III) solution in ml, m is weight of the biomass in g.

The percentage of removal of Cr (III) ions ($R\%$) in aqueous solution was calculated using Eq. 2:

$$R(\%) = \frac{(C_i - C_f)}{C_i} \times 100 \quad (2)$$

RESULTS

Characterization of Microbial Biomass and Tannery Effluent

Different physico-chemical parameters of *Aspergillus* biomass (MB1) are as follows: pH, 5.0 ± 0.52 ; bulk density (g cm^{-3}), 0.58 ± 0.35 ; ash (%), 5.58 ± 0.56 ; moisture (%), 7.50 ± 0.52 ; carbon (%), 46.24 ± 0.20 ; hydrogen (%), 6.8 ± 0.42 ; nitrogen (%), 3.22 ± 0.54 ; calorific value (MJ kg^{-1}), 16.42 ± 1.51 . Tannery effluent was acidic (pH, 3.5), greenish in colour with Cr (III) (1700.9 mg L^{-1}), total solids ($96,000 \text{ mg L}^{-1}$), total dissolved solids ($80,000 \text{ mg L}^{-1}$) had Chemical Oxygen Demand (COD) of 332.8 mg L^{-1} and Biochemical Oxygen Demand (BOD) of 290 mg L^{-1} . Traces of other heavy metals such as Fe, Pb Co, Cu, Cd and secondary elements like Ca, Mg and Na were also present.

Surface Characterization of Microbial Biomass

To determine the behaviour of microbial biomass towards adsorption of Cr (III), X-Ray Diffraction (XRD) and Fourier infrared spectroscopy (FTIR) of native and tannery effluent loaded biomass were carried out. X-Ray diffraction of both biomass before and after saturation with Cr (III) was carried out to identify the phase present on adsorbent (Fig. 1, 2). The XRD analysis of *Aspergillus* sp. biomass did not show any peak after tannery effluent treatment indicating amorphous nature of the biomass. The FTIR spectrum of biomass displayed a number of adsorption peaks at 1655, 1546, 1319 and 1034 cm^{-1} indicating the presence of CN, CC, NH, CS-NH, OH, C-Cl, C-O groups on the biomass surface Fig. 3.

Batch Sorption Studies

Batch sorption studies on Cr (III) removal from aqueous solution by *Aspergillus* sp. (MB1) was carried out to optimize various parameters such as adsorbent dosage, pH, chromium concentration and contact time, removal of Cr (III) from tannery effluent.

Effect of Adsorbent Dosage

Effect of adsorbent dosage (0.25-2%) on sorption of Cr (III) from aqueous solution (pH4) containing (25 mg L^{-1}) with agitation rate of 120 rpm at ambient temperature was studied. Increasing adsorbent dosage (0.25-2%), Cr (III) removal increased rapidly from 24.3 to 94.6% by (*Aspergillus* sp. (MB1)).

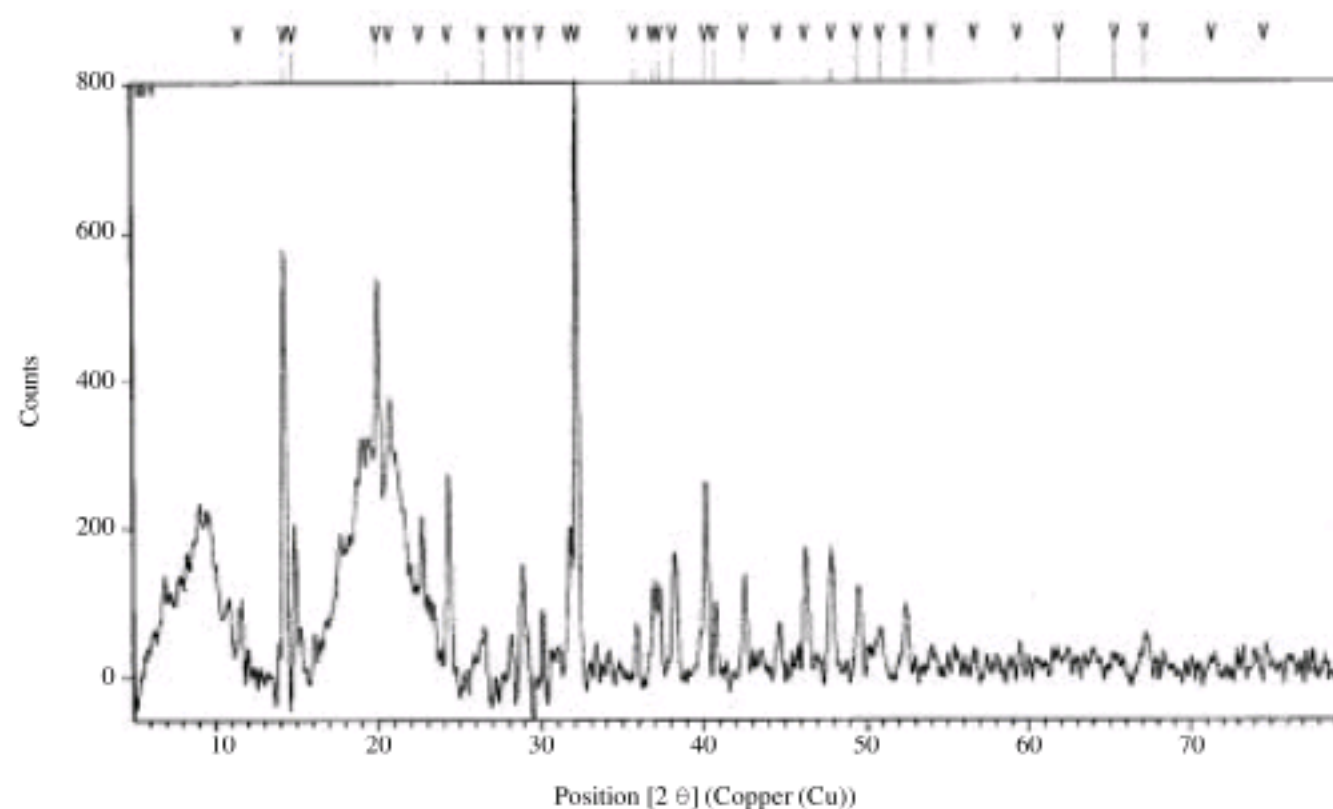


Fig. 1: XRD patterns of *Aspergillus* sp. (MB1) biomass before treatment

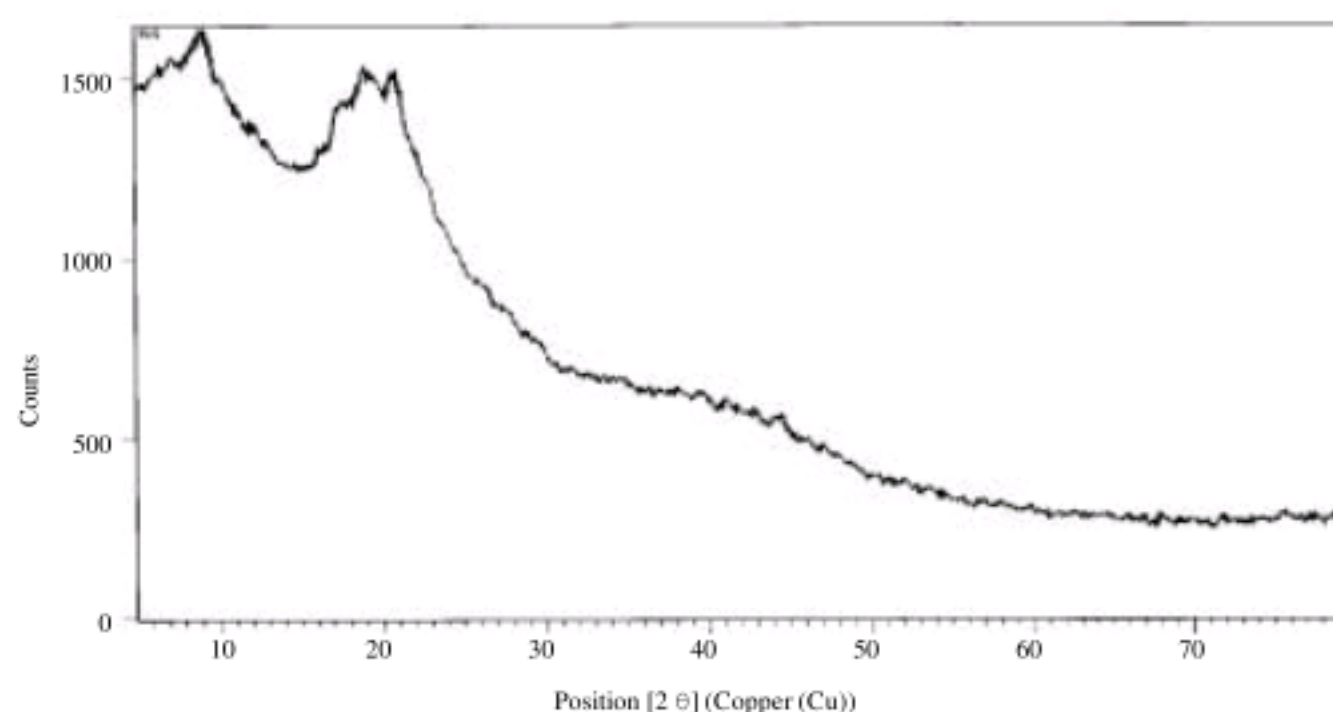


Fig. 2: XRD patterns of *Aspergillus* sp. (MB1) biomass after tannery effluent treatment

Effect of pH

The amount of Cr (III) adsorbed increased from 1.17 to 2.24 mg g⁻¹ in *Aspergillus* sp. MB1 as the pH increased from 2 to 5. Maximum Cr (III) removal of 94.6% was observed with *Aspergillus* sp. (MB1) at pH 4.0.

Effect of Chromium Concentration

Kinetic studies of Cr (III) removal by *Aspergillus* sp. biomass (MB1) was conducted at varying initial concentrations (5-50 mg L⁻¹) of Cr (III). Maximum removal of Cr (III) from aqueous solution and tannery effluent by *Aspergillus* sp. biomass was 94.6 and 64.48%, respectively. Removal of Cr (III) was dependent on concentration of chromium because increase in initial chromium concentration.

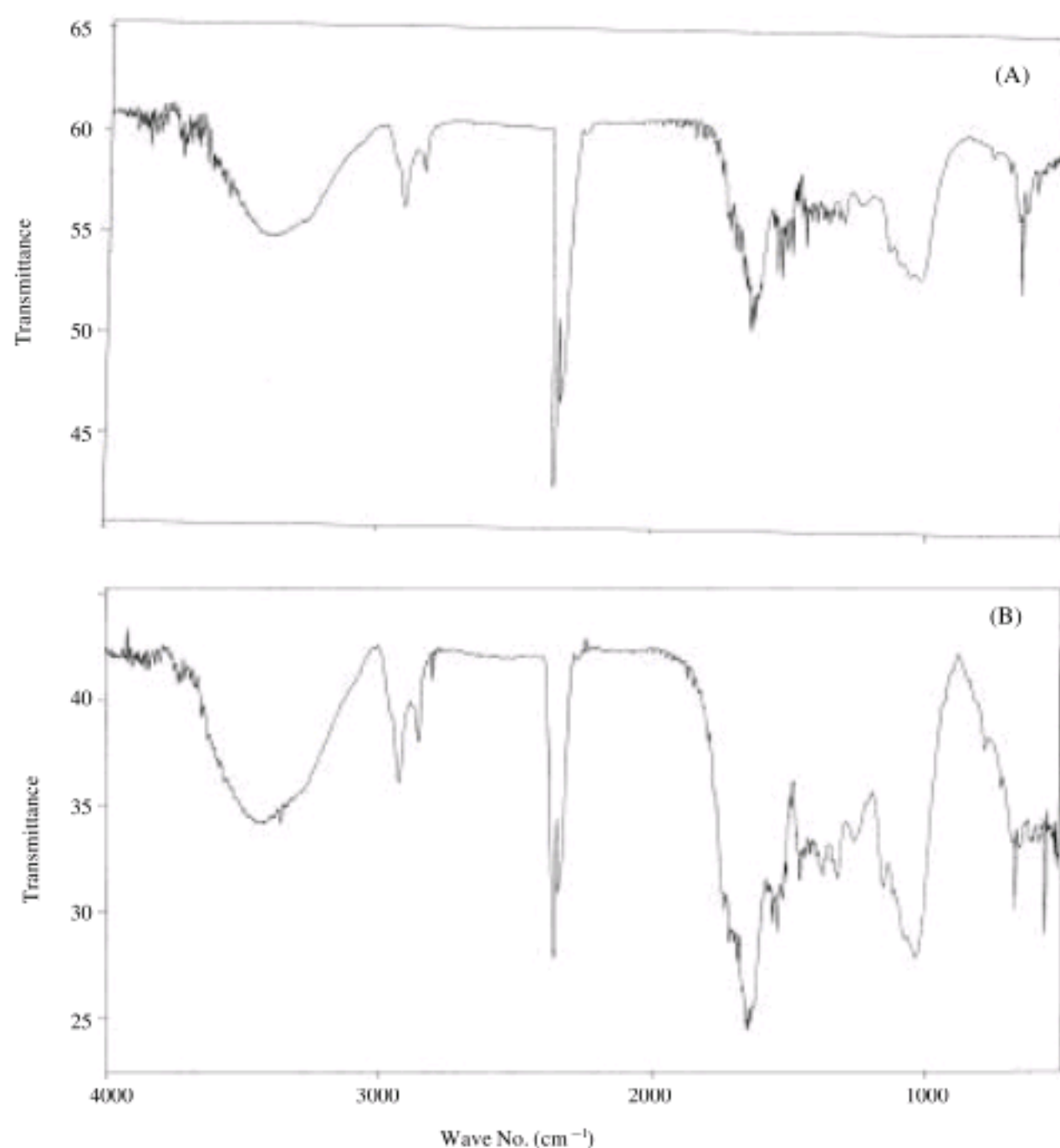


Fig. 3: FTIR spectra of *Aspergillus* sp. (MB1) biomass, (A) untreated biomass and (B) tannery effluent treated biomass

Effect of Contact Time

Biosorption of chromium onto *Aspergillus* sp. biomass increased with contact time from 0.08 to 4 h led to an increase in Cr (III) removal from 40-70% and 20-50% of Cr (III) from aqueous solution tannery effluent, respectively within first 2 h, which represents the time at which equilibrium of chromium biosorption is presumed to have been attained. This suggests that the adsorption is rapid and typically 40-50% of the ultimate adsorption occurs within the first hour of contact and saturation is reached with in next 48 h.

Adsorption Kinetics

Adsorption kinetics of Cr (III) study describes solute uptake rate controls which residence time of adsorbate uptake at solid and solution interface. Conformity between experimental data and the models predicted values was expressed by correlation coefficients (R^2 values near or equal to 1). A relatively high R^2 value indicates that model successfully described kinetics of Cr (III) adsorption.

The Pseudo First-Order Equation

Lagergren (1989), the Pseudo first-order equation is generally expressed as follows:

Table 1: The adsorption kinetic model rate constants for the MB1 (*Aspergillus* sp.) at different adsorbent dosage ($C_i = 50 \text{ mg L}^{-1}$, adsorbent dosage 1%, pH 4, 120 rpm)

Initial pH	Pseudo first-order			Pseudo second-order			Elovich model		
	k_1	q_e	r^2	k_2	h	r^2	β	α	r^2
0.25	0.03	8.64	0.84	2.29	2.44	0.98	3.21	21.29	0.83
0.5	0.02	3.84	0.79	1.80	2.61	0.98	3.24	9.67	0.93
1	0.06	0.43	0.75	2.14	20.00	0.99	19.61	0.40	0.92
2	0.05	0.21	0.96	1.12	5.07	0.99	33.44	0.09	0.93

Table 2: The adsorption kinetic model rate constants for the MB1 (*Aspergillus* sp.) at different pH ($C_i = 50 \text{ mg L}^{-1}$, pH 4, 120 rpm)

Adsorbent dose (g)	Pseudo first-order			Pseudo second-order			Elovich model		
	k_1	q_e	r^2	k_2	h	r^2	β	α	r^2
2	2.00	0.02	1.57	1.09	5.95	0.99	10.67	13.90	0.69
3	0.01	1.95	0.72	1.04	1.58	0.99	10.56	80.60	0.29
4	0.76	3.03	0.88	2.43	31.35	1.00	21.69	0.18	0.90
5	0.03	3.29	0.71	2.24	84.03	1.00	55.56	0.13	0.75

Table 3: The adsorption kinetic model rate constants for the MB1 (*Aspergillus* sp.) at different concentration of chromium (Adsorbent dosage 1%, pH 4, 120 rpm)

Cr (III) (mg L^{-1})	Pseudo first-order			Pseudo second-order			Elovich model		
	k_1	q_e	r^2	k_2	h	r^2	β	α	r^2
5	0.03	0.46	0.68	0.32	0.22	0.97	19.16	0.90	0.89
10	0.04	0.88	0.82	0.48	0.63	1.00	14.49	3.65	0.95
25	0.08	1.92	0.98	1.56	0.41	0.98	4.01	1.62	0.91
50	0.05	3.89	0.93	2.77	3.84	1.00	2.92	0.61	0.91

$$\frac{dq_t}{dt} = k_1(q_e - q_t) \quad (3)$$

where, q_e and q_t are the adsorption capacity at equilibrium and at time t , respectively (mg g^{-1}), k_1 rate constant of pseudo first-order adsorption process (L min^{-1}). After integration and applying initial conditions $t = 0$ to $t = t$ and $q_t = 0$ to $q_t = q_t$, the integrated form of Eq. 3 becomes:

$$\log(q_e - q_t) = \log(q_e) - \frac{k_1}{2.303} t \quad (4)$$

Plot of $\log(q_e - q_t)$ vs. t should give a linear relation from which k_1 and q_e can be determined from slope and intercept of the plot, respectively. R^2 of Cr (III) under different conditions were calculated from plots at different pH (Table 1), biomass dosage (Table 2) and Cr (III) concentration (Table 3). A plot of $\log(q_e - q_t)$ versus agitation time, gave a straight line with R^2 of 0.88 for MB1 at pH 4 (Table 1-3). Results showed that adsorption reaction can be approximated with pseudo first-order kinetic model. It was observed that k_1 (0.03-0.05) and q_e (8.64-0.21) at different adsorbent dosage (0.25-2%) and values of k_1 (2.00-0.03) and q_e (0.02-3.29) at different pH. In most of the cases, first-order equation of Lagergren did not apply throughout complete range of contact time and is generally applicable over initial (20-30 min) sorption process. The pseudo second-order equation (Ho and Mckay, 1999; Ho *et al.*, 1996).

The pseudo second-order adsorption kinetics model based on the sorption capacity of solid phase can be used in this case assuming that measured concentrations are equal to the cell surface concentrations. If the rate of sorption is a second order mechanism, the pseudo second-order chemisorption kinetic equation for Cr (III) both singly and in the mixture is expressed as:

$$\frac{dq_t}{dt} = k_2(q_e - q_t)^2 \quad (5)$$

where, k_2 is rate constant of pseudo second-order adsorption (g/mg/min). For initial condition $t = 0$ to $t = t$ and $q_t = 0$ to $q_t = q_e$, integrated form of Eq. 5 becomes:

$$\frac{1}{(q_e - q_t)} = \frac{1}{q_e} + kt \quad (6)$$

This is integrated rate law for a pseudo second-order reaction. Equation 6 can be rearranged as, which has a linear form:

$$\left(\frac{t}{q_t} \right) = \frac{1}{k_2 q_e^2} + \frac{1}{q_e} (t) \quad (7)$$

If the initial adsorption rate, h (mg/g/min) is:

$$h = k_2 q_e^2 \quad (8)$$

Then Eq. 7 and 8 becomes:

$$\left(\frac{t}{q_t} \right) = \frac{1}{h} + \frac{1}{q_e} (t) \quad (9)$$

Plot of (t/q_t) and t of Eq. 7 should give a linear relationship, from which q_e and k_2 can be determined from slope and intercept of plots, respectively. R^2 for linear plots are found superior (in most cases = 0.99) (Table 1-3). It was observed that k_2 (2.29-1.12) and h (2.44-5.07) increase with increase in biomass dosage and k_2 Values of k_2 (1.09-2.24) and h (5.95-84.03) increased at different pH. Values of k_2 (0.32-2.77) and h (0.22-3.84) increased with increase in initial Cr (III) concentration. Values of h and k_2 were found higher for MB1 at different parameters (adsorbent dosage, pH and Cr(III) concentration). Experimental points shown together with theoretically generated values reflect extremely high correlation coefficients (Table 1-3).

The Elovich Equation

The Elovich model equation is generally expressed as (Chien and Clayton, 1980):

$$\frac{dq_t}{dt} = \alpha \exp(-\beta q_t) \quad (10)$$

where, α is the initial adsorption rate (mg/g/min), β is the desorption constant ($g \text{ mg}^{-1}$) during any one experiment. To simplify the Elovich equation, assumed $\alpha\beta t \gg 1$ and by applying the initial conditions ($q_t = 0$ at $t = 0$ and $q_t = q_t$ at $t = t$) Eq. 10 becomes:

$$qt = \frac{1}{\beta} \ln(\alpha\beta) + \ln(t) \quad (11)$$

If, Cr (III) adsorption fits Elovich model, a plot of qt vs. $\ln(t)$ should yield a linear relationship with a slope of $(1/\beta)$ and an intercept of $(1/\beta) \ln(\alpha\beta)$. R^2 were determined from plot between q_t and $\ln(t)$, respectively at different biomass dosage, pH and chromium concentration (Table 1-3). Computed value of α and β were calculated at biomass dosage [α (21.29-0.09), β (3.21-33.44)],

pH [α (13.90-0.13) β (10.67-55.56)] and Cr (III) concentration [α (0.90-0.61) β (19.16-2.92)]. Sorption kinetics of Cr (III) examined using Elovich model gave R^2 of 0.90 for MB1, which indicates that dynamics data fitted for biomass at pH 4.

DISCUSSION

Physicochemical treatment of such effluent is not only expensive but not eco-friendly. Use of low cost natural resource as the adsorbents for the removal of chromium is constantly encouraging in preference to other conventional methods. Biosorption offers excellent potential for the removal of metals from effluent. A multitude of biomass types comprising fungal, bacterial, algae, peat biomass and so on have been studied for metal biosorption (Kapoor and Viraraghavan, 1998; Chandrasekhar *et al.*, 1998; Tobin and Roux, 1998; Gupta *et al.*, 2001; Cossich *et al.*, 2004). Different non conventional biosorbents of plant and microbial origin were tested for removal of heavy metals from synthetic effluents in single metal and multi-metallic state of industrial effluent to develop an inexpensive adsorbent. The process can be made economical by procuring natural bulk biomass or spent biomass from various fermentation industries (Ahluwalia and Goyal, 2007). Waste biomass generated by different industrial processes was developed as a biosorbent for its large scale application for removal of chromium from mixed metal effluent (Gupta *et al.*, 1999; Dias *et al.*, 2000; Li *et al.*, 2004; Fahim *et al.*, 2006).

In this study, greenish tannery effluent (pH, 3.5; high Cr (III), 1700.9 mg L⁻¹) was treated with microbial waste biomass from fermentation industry. Physicochemical parameters of *Aspergillus* sp. biomass and calorific value were found to coincide with the relative calorific value (17.5 MJ kg⁻¹) of solid fuels such as biomass, municipal waste, industrial waste, peat and brown coal (MacDonald *et al.*, 1996). The XRD data of treated adsorbent provided evidence of decrease in the peak intensity which shows the adsorption of Cr(III) on the surface of the adsorbent changing its structure from crystalline to amorphous. This is in agreement with the previous studies on adsorption (Gopal and Elango, 2007; Karthikeyan and Ilango, 2008).

Comparison of the FTIR spectra of native and tannery effluent treated biomass revealed that new adsorption peaks at 2360, 1451 and 1376 cm⁻¹ were formed indicating the involvement of amine N-N=O, CS-NH, OH group, respectively. The assignment of a specific wave number to a given functional group was not possible because the adsorption bands of various functional groups overlap and shift, depending on their molecular structure and environment. Shifts in absorption positions may be caused by many factors. These include: (1) the physical state (2) electronic and mass effects of neighbouring substitutes (3) conjugation (4) intra-molecular and intermolecular hydrogen bonding and (5) ring strain (Mohan *et al.*, 2006).

In batch sorption studies chromium removal affinity of the microbial biomass was determined by aqueous solution and tannery effluent at different adsorbent dosage (0.25-2 g), pH (2-5), initial chromium (III) concentration (5-50 mg L⁻¹) and contact time. With increasing adsorbent dosage (0.25-2%), Cr (III) removal increased rapidly from 24.3 to 94.6% by (*Aspergillus* sp. Similar observation were made with *Aspergillus* sp. (Chandrasekhar *et al.*, 1998) which showed 70-75% removal of Cr (III) from aqueous solution and exhibited an adsorption capacity of 1.8-2 mg g⁻¹ (Mattuschka and Straube, 1993) and comparable with commercial activated carbon (2.7 mg g⁻¹). Chromium exhibits different types of pH dependent adsorption (Mohan *et al.*, 2006) maximum (94.6%) removal of Cr (III) was observed with *Aspergillus* sp. (MB1) at pH 4.0. Similar observations were made for biowaste generated as a by-product of large-scale industrial fermentation (Dias *et al.*, 2000; Cossich *et al.*, 2004; Fahim *et al.*, 2006; Chandrasekhar *et al.*, 1998). It is clear that sorption of

trivalent chromium species increased with increase in solution pH from 2 to 5.0. Studies were not performed beyond pH 5.0 due to the possibility of Cr (III) precipitation (Mohan *et al.*, 2006). Adsorption of Cr (III) is therefore an electrostatic interaction between biomass and Cr (III) cations. Initially, the immediate solute uptake achieved within a 2 to 4 h, subsequently the second stage of solute uptake continued for longer time (24 h) period (Demirbas *et al.*, 2004).

Maximum removal of Cr (III) from aqueous solution and tannery effluent by *Aspergillus* sp. biomass was 94.6 and 64.48%, respectively. Earlier reports also indicate that removal of Cr (III) is dependent on concentration of chromium because increase in initial chromium concentration (50-300 mg L⁻¹) increases the amount of Cr (III) adsorbed (Demirbas *et al.*, 2004).

Biosorption of chromium onto *Aspergillus* sp. biomass increased with contact time from 40-70% and 20-50% of Cr (III) from aqueous solution tannery effluent, respectively. Diminishing removal with increasing time may be due to intraparticle diffusion process dominating over adsorption (Chandrasekhar *et al.*, 1998; Kapoor and Viraraghavan, 1998).

The kinetics of Cr (III) adsorption on the *Aspergillus* sp. were analyzed using pseudo first-order (Lagergren, 1989), pseudo second-order (Ho *et al.*, 2000), Elovich (Chien and Clayton, 1980) and intraparticle diffusion (Srivastava *et al.*, 1997) kinetic models. A relatively high R² value indicates that the model successfully describes the kinetics of Cr (III) adsorption. Adsorption kinetic of Cr (III) removal has been further studied at different adsorbent dosage, pH, chromium concentration and contact time. In most of the cases, first-order equation of Lagergren did not apply throughout the contact time and is generally applicable over initial (20-30 min) sorption (Mohan *et al.*, 2006; Lagergren, 1989). Slope and intercepts as calculated from plots were used to determine k₁ (first-order rate constant) and equilibrium capacity (q_e). Values of q_e are found lower than experimental one. Therefore, chromium-adsorbent systems do not follow a first-order rate equation (Mohan *et al.*, 2006; Ho, 2004; Demirbas *et al.*, 2004). Data showed good compliance with pseudo-second order kinetic model (R²>0.99) (Mohan *et al.*, 2006; Fahim *et al.*, 2006; Li *et al.*, 2004). However, these two models do not provide a definite mechanism. Therefore, another simplified model (Ho *et al.*, 1996; Ho and McKay, 1999) was tested, in which kinetic data is better fitted by second-order rate equation (Mohan *et al.*, 2006; Ho, 2004; Demirbas *et al.*, 2004). Experimental q_e values were compared with q_e value determined by pseudo-first and second-order rate kinetic models, which suggests that sorption system is not a pseudo first-order reaction and that a pseudo-second-order model can be considered (Demirbas *et al.*, 2004).

Novelty and significant contribution of the present study is that the capability of industrial waste microbial biomass (*Aspergillus* sp.) for the removal of trivalent chromium from aqueous solution and tannery effluent has been studied. The waste biomass can be easily procured from industrial fermentation processes as byproduct and no cost is involved in its production for using as sorbent which serve as highly economical with constant supply in an eco-friendly manner. Present investigation revealed that the industrial waste biomass at lower concentration can efficiently remove Cr (III) under various laboratory conditions. In addition, kinetic models were in agreement with experimental data, which clearly indicates its significant role in remediation of chromium (III) and leading to a product as well as process development which will have industrial application and commercial value and significance in pollution abatement. The strategy will lead to value addition to the waste biomass of one industry and the product can be used for treatment of effluent from another industry which will be one time treatment of metal containing effluent leading to its detoxification in cost effective and eco-friendly manner. Present study provides a significant contribution towards the trivalent chromium removal by waste biomass, which can be subsequently extrapolated for treatment of industrial wastewater.

CONCLUSION

Microbial waste biomass comprising of *Aspergillus* sp. involved in fermentation process was effective in removal of Cr (III) from aqueous solution as well as tannery effluent. To our knowledge, this microbial biomass does not find any economic application and represents a solid pollutant to the environment. In this study, maximum chromium removal capacities of *Aspergillus* sp. biomass (MB1) were 94.6% from aqueous solution and 64.48% from tannery effluent at ambient temperature, rpm 120 and initial concentration (50 mg L^{-1}) of Cr (III). A comparison of kinetic models (Lagergren, Ho and McKay, Elovich and Morris-Weber) applied to MB1 indicates that adsorption of Cr (III) on MB1 follows a best Ho and McKay pseudo second-order rate equation and correlation coefficient (R^2) correlated with experimental data. XRD and FTIR studies reveals that Cr (III) adsorption takes place on the surface of the adsorbent. This can provide means for using solid waste of one industry to treat chromium containing effluents from another industry for development of a suitable cost effective biosorbent.

ACKNOWLEDGMENTS

The authors are thankful to Council of Scientific and Industrial Research (CSIR) New Delhi, vide Sanction No. 24 (0284)/05/EMR-II, for providing financial support, Ranbaxy Paonta Sahib, Himachal Pradesh, India for providing microbial biomass and director, Thapar University for providing infrastructural support.

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