



Research Journal of  
**Environmental  
Toxicology**

ISSN 1819-3420



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## Determination of the Effect of Red Earth as an Ameliorant for Lead Contaminated Soils\*

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**Abstract:** The aim of this study was to assess the potential of Red earth/mud as an ameliorant for Pb contamination. The results showed a significant reduction of toxicity with soils contaminated with PbS, PbSO<sub>4</sub>, PbNO<sub>3</sub> and PbCO<sub>3</sub>. Critical issues raised during the study showed that red mud drastically reduced lead contamination most probably due to shifting of lead from the exchangeable to the Fe-oxide fraction through specific chemisorption and diffusion mechanisms including a reduction of solubility and mobility of Pb. However, speciation pattern of different lead compounds need to be tied to bioavailability to reliably characterize the adsorption behaviour of lead.

**Key words:** Red earth/mud, toxicity, ameliorant, lead contamination

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### INTRODUCTION

Red earth, also known as red mud, is a waste material formed through the Bayer process of alumina production in the aluminum industries (Lopez *et al.*, 1998). Red earth contains mainly iron oxide (30-50%), with significant amounts of silica, aluminum calcium oxide and titanium, dispersed in a highly alkaline and caustic liquor (Nguyen and Boger, 1998). The treatment and disposal of this bauxite residue is a major operation and may account for 30-50% of the operational cost in an alumina refinery (Nguyen and Boger, 1998). According to Wong and Ho (1995), the red earth has been accumulated at a rate of over 60 million tonnes annually throughout the world. The toxicity and colloidal nature of the red mud particles as well as the relatively large quantities generated create a serious pollution hazard and necessitate improved disposal techniques (Akay *et al.*, 1998). There have been many proposals of red mud utilization and currently, its use is in the manufacture of building materials and ceramics, as a filler in asphalt road, as an iron ore and a source of various minerals, but still a large volume is dumped in holding ponds for which a large area of land is required (Li, 1998).

Friessl *et al.* (2003) assessed the effects of various soil amendments on the lability of metals in relation to uptake by plants, dry-matter yield, pH and overall bioavailability. The results of a sequential extraction procedure revealed a relative enrichment of Cd in the Fe-oxide-associated fraction from red mud amended soil, suggesting an additional immobilization mechanism along with pH increase (Lombi *et al.*, 2002). The same workers also found that of all the amendments tested, red mud was the most effective for increasing soil pH. This increase was most pronounced for acidic soil. In general, red mud-amended soils maintained pH values between 7.1 and 7.9 throughout their experiment, while the other amendments had little or no effect on soil pH.

Red earth was also used for the removal of nickel toxicity by Zouboulis and Kydros (1993), where red earth was found to act simultaneously as an alkalinity regulator. Lombi *et al.* (2002) assessed the ability of red mud to reduce mobility and availability of a range of heavy metals in soils

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\*Originally Published in Research Journal of Environmental Toxicology, 2007

contaminated by different sources and the results suggested that specific chemisorption and possibly metal diffusion into oxide particles, could also contribute to the mechanisms responsible for the fixation of metals by red mud.

Friessl *et al.* (2004) assessed the effects of red mud on metal lability and bioavailability in soil and concluded that application rates have to be adjusted according to soil conditions (e.g., pH). In some cases, removal of pollutants (e.g., As, Cr, V) associated with red mud using leaching procedures may be required. Qiao and Ho (1999) investigated the effect of clay amendment on speciation of heavy metals in sewage sludge and found that red mud addition to sewage sludge compost significantly affected metal specification. Red mud reduced metal mobility and the potential hazard of releasing metal from sludge due to further breakdown of organic compounds.

The adsorption of toxic heavy metal cations, such as Cu(II), Cd(II) and Pb(II), by red mud has been modelled (Apak *et al.*, 1998) with the aid of a modified, Langmuir equation and a double-site binding model incorporating the effect of pH. It was concluded that the adsorption of the heavy metal cations, Cu(II), Cd(II) and Pb(II), by red earth can be interpreted with respect to pH and metal concentration by means of this approach. Heavy metal removal from aqueous and soil solutions utilising red mud has been studied and saturation capacities of red mud for Cu(II), Cd (II) and Pb (II) were found to be 63, 105 and 123 mg g<sup>-1</sup>, respectively (Apak *et al.*, 1993; 1998; Apak, 1991). Red mud, as a composite adsorbents consisting of various hydrated oxides, showed a high capacity for these metals, necessitating further investigation of their binding potential onto the adsorbent surface.

With a review to simultaneously conserve the environment and utilising an important waste resource, Lopez *et al.* (1998) assessed the feasibility of using residues from bauxite refining, for the purpose of wastewater treatment. This experiment suggested that, after pre-treatment to neutralize alkaline components and induce formation of stable aggregates, red mud was suitable for use as a filtration medium for treatment of wastewaters, particularly those whose principal contaminants were phosphorus or heavy metals.

Increasingly stringent legislation regarding the quality of water has created a growing interest in the improvement of conventional treatment processes. Various methods of wastewater treatment have been examined and amendments based on adsorption have emerged as some of the most promising techniques (Cheremisnoff and Ellerbush, 1979; Pollard *et al.*, 1992). Such techniques have included the use of activated carbon which is still very popular with different grades available, although it is quite expensive and the complete regeneration of carbon is not always possible (Lalvani *et al.*, 1998). Considerable research work has been carried out to explore inexpensive adsorbents, especially industrial waste materials such as fly ash (Ferraiolo *et al.*, 1990; Diama-doloulos *et al.*, 1993), metal hydroxides (Namasi-vayam and Ranganathan, 1998), blast furnace slag (Dimitrova and Mehasndgiev, 1998), biomass (Chang *et al.*, 1997) peanut hull (Persiansamy and Namasivayam, 1994), bagasse pith (Aly and Daifullah, 1998), carbonaceous material (Srivasta *et al.*, 1989, 1997; Srivasta and Tyagi, 1995) and bagasse fly ash (Park *et al.*, 1999).

Gupta *et al.* (1997, 1998 and 1999a) investigated the possible utilization of solid waste materials, being generated in some prime industries, for immobilisation of metals. Particular focus was applied to the potential of red mud as an adsorbent for the removal of toxic metal ions (e.g., lead and chromium) from aqueous solution (Gupta *et al.*, 1998). It was concluded that red mud obtained from the aluminum industry was an efficient source for the removal of lead and chromium and could be used as a suitable adsorbent for the treatment of wastewater of various industries. Red earth/mud is also known to be suitable for removal of chlorophenols from water. Altundogan *et al.* (2000) used red earth to remove arsenic from aqueous solution by adsorption and the test showed that the resulting alkaline aqueous medium (pH 9.5) favoured the removal of As(III), whereas lower pH's required for the removal of As(V). However, the characteristic of Red earth in aspects such as its binding mechanisms, influence at pH of spiked soils and adsorption potential of red mud needs further investigation during this study.

The aim of this study was to assess the potential of Red earth/mud as an ameliorant for Pb contamination.

## **MATERIALS AND METHODS**

The topsoil used in these experiments was collected from the top 25 cm of an agricultural field on the Craibstone estate, which is approximately 9 km North West of Aberdeen, at an elevation of 100 m. The pots had a diameter of 150 mm and a surface area of 17,600 mm<sup>2</sup>. A Whatman 42 filter paper was placed on the base of each pot to prevent coarse material from passing through. Leaching pots were arranged on a leaching bench with holes wide enough to hold them. Funnels with aligned filter paper (Whatman 42) inside were placed under each pot placed on the leaching bench to collect leachate in a conical flask placed on the bottom of the shelf.

### **Preparation of Experimental Pots**

Thirty pots were packed with either a mixture of soil, lead compounds and Red earth, or soil with lead compound only.

Control: 3 pots of soil only  
          3 pots of soil + Red earth  
Samples: 3 pots of soil + Pb compound  
          3 pots of soil + Pb compound + Red earth

Since 4 Pb compounds were considered (PbS, PbSO<sub>4</sub>, PbNO<sub>3</sub>, PbCO<sub>3</sub>), the total number of samples was 24. This translates to thirty pots together with 6 pots of control.

The mixture of soil, Pb compound and Red earth under study was thoroughly shaken together in plastic bags of 1 kg capacity to allow homogeneity prior to packing the pots. The amount of red earth added to 1 kg soil in a pot, was 6.92 g.

The area where the bottles were placed was protected with black, plastic material to minimize the effect of light on leachate chemical properties. All the experimental treatments were carried out in triplicate.

### **Bioassay**

*Lux*-marked bacterial biosensors were used during the study and the preparation of the biosensor and luminometer measurements were carried out as described. One hundred microliter of the resuscitated biosensor suspension was added to the samples at 15 sec intervals, accurately timed for measurement in the Bio Orbit 1253 luminometer (Labtech International, Uckfield, U.K). Each sample was exposed to the sensor for exactly the same time. Samples were incubated for 15 min before light output measurements were carried out at 15 sec intervals. This ensured the same exposure time to the potentially toxic elements for cells in each of the cuvettes.

### **Chemical Analysis**

#### **Stock Solution Preparation**

1.599 g of lead nitrate, Pb(NO<sub>3</sub>)<sub>2</sub> (analytical grade) was carefully weighed and dissolved in deionized distilled water. When dissolution was complete, it was acidified with 1 mL of 1M HNO<sub>3</sub> and diluted to 1 L with deionized water.

#### **Preparation of Standard Solutions and Calibration**

Standard Lead Solution was prepared by diluting the stock (lead) solution. Concentration ranges starting from 0.1, 0.5, 100, 200, 400, to a maximum of 800 g L<sup>-1</sup>) which were used as calibration standards. Standard solutions of lead were prepared fresh for use from a stock solution of lead nitrate

(0.1 mol L<sup>-1</sup> in HNO<sub>3</sub>). All standard and sample soil solutions were prepared to approximately 0.1 mol L<sup>-1</sup> in HNO<sub>3</sub>. Care was taken to use specially purified water (deionized water) when diluting samples to final volume for quality control purposes.

Deionized water was also used during the final rinsing of all the plastic and glassware. This was after rinsing them first, in solution (with diluted nitric acid) in order to remove any possible traces of lead on them. During the determination of concentration two replicate determinations of absorbance were made for each sample. A blank of dionized water was used to zero the instrument.

A 10 µL sample was injected very carefully with the help of an auto sampler into the cold graphite furnace and, by means of an automatic temperature programmer, dried at 120°C for 35 sec and at 140°C for another 35 sec, then heated to 200°C and allowed to cool for 15 sec. These steps were performed, automatically, to remove solvent and any removable volatile matrix. Actual atomization of the sample followed and was performed at 1800°C, very rapidly, for 5 sec. During this time the signal from the chamber (absorbance) was recorded and displayed on the screen as a function of time. Finally the furnace was heated for 5 sec at 2600°C. The purpose was to remove any residues and prepare the instrument for next sampling phase. During the atomization step, the absorbance was monitored at 283.3 nm, using a slit width of 0.7 nm, set at low level. Purging with argon was interrupted automatically during the absorbance scan. Background correction was provided by means of the deuterium background corrector, which automatically compensated for broadband absorption interferences.

#### **Data Analysis**

Two-way analyses of (ANOVA-Analysis of Variance) (except for biosensor experimental data which is One-way ANOVA) were carried out using the statistical package Minitab for windows, release 12.1 (State College, PA, USA). Mean differences were determined using t-test (paired two samples for means) and Pearson Correlations using Excel program (Microsoft™ Office 2000). Significant differences between treatments were elucidated using least significance difference (LSD) values. Graphs were generated using SigmaPlot for Windows version 9.0 (Jandel Corporation, CA and USA).

## **RESULTS**

### **Chemical Analysis**

#### **Effect of Red Earth on Lead Concentration of Leachate**

Red earth showed a significant effect as an ameliorant (i.e., reducing leaching of Pb) for soils contaminated with PbS and PbCO<sub>3</sub> (p<0.001), PbSO<sub>4</sub> and PbNO<sub>3</sub> (p<0.01) over a period of 230 days as shown in Fig. 1 a-e. Similarly, the effect of red earth as an ameliorant was significant over time (days) and this was only observed for PbS (p<0.01) and PbNO<sub>3</sub> (p<0.05). This observation was considered important in determining the suitability of red earth as an ameliorant of lead contaminated soils. Table 1 showed the significance of time factor when lead compounds were treated with Red earth/mud. All the treatments were affected by time as a factor except for PbSO<sub>4</sub> and PbCO<sub>3</sub>. The time factor for contaminated soils treated with Red earth/mud is critical for design of mitigation strategies

**Table 1: Comparison of the significance of the effect of bone meal treatment and time**

(Source of Pb)	Treatment (Bone meal)	Time factor
PbS	***	*
PbSO <sub>4</sub>	**	ns
PbNO <sub>3</sub>	**	*
PbCO <sub>3</sub>	***	ns

\*\*\* p<0.001; \*\* p<0.01; \* p<0.05, ns: Non Significant

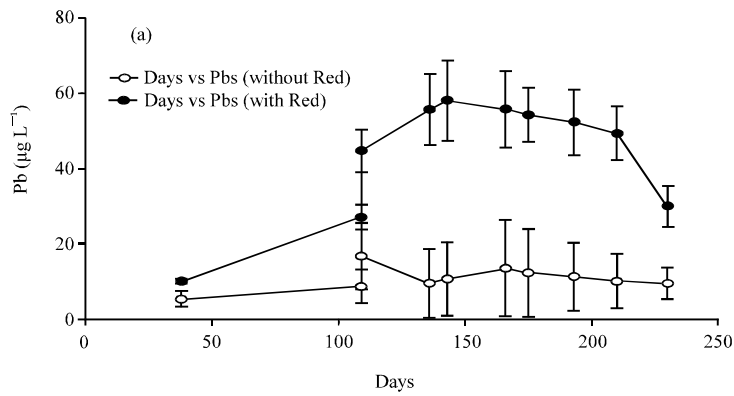


Fig. 1a: Lead concentration ( $\mu\text{g L}^{-1}$ ) in leachate from soil contaminated with (a) PbS and treated/untreated with Red earth over a period of 230 days

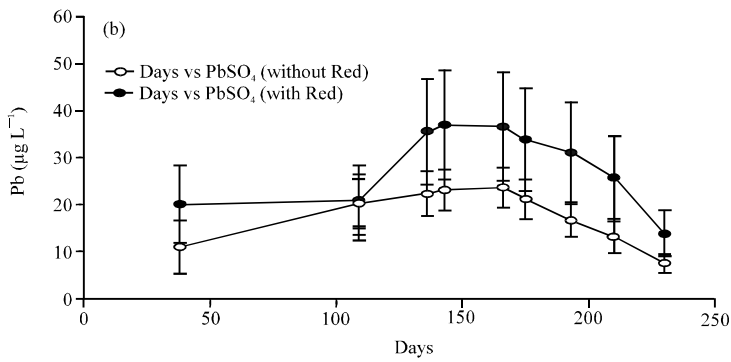


Fig. 1b: PbSO<sub>4</sub> and treated/untreated with Red earth over a period of 230 days

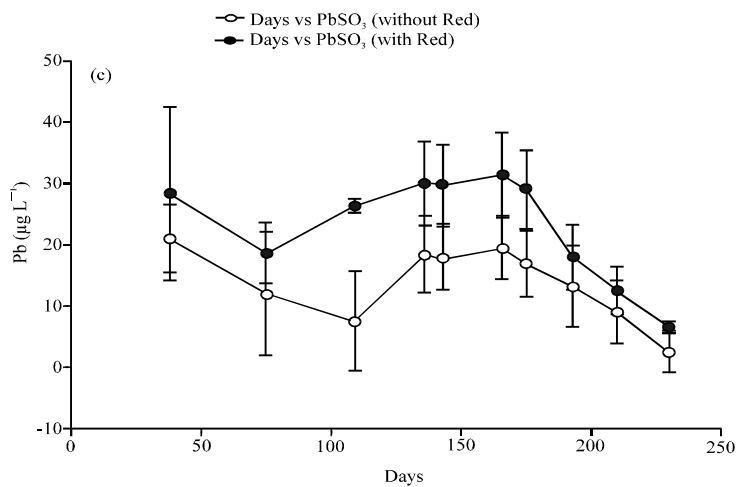


Fig. 1c: PbNO<sub>3</sub> and treated/untreated with Red earth over a period of 230 days

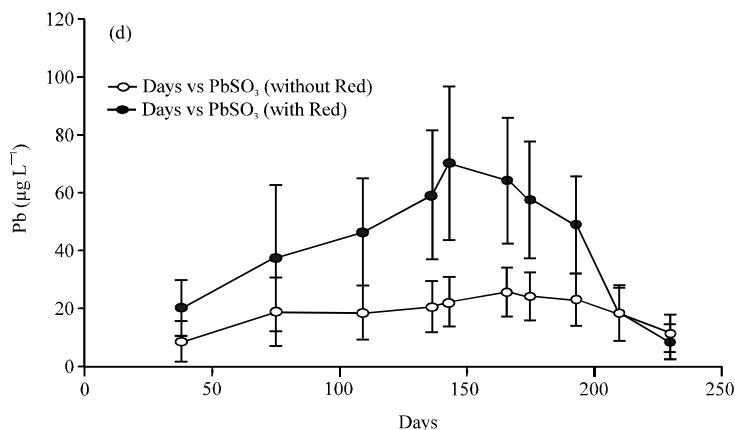


Fig. 1d: PbCO<sub>3</sub> and treated/untreated with Red earth over a period of 230 days

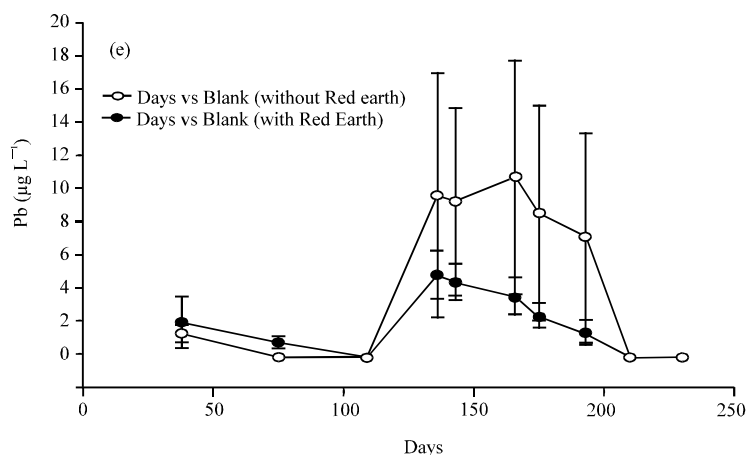


Fig. 1e: Lead concentration ( $\mu\text{g L}^{-1}$ ) in leachate from soil contaminated with (i) PbS, (ii) PbSO<sub>4</sub>, (iii) PbNO<sub>3</sub>, (iv) PbCO<sub>3</sub> and treated/untreated with Red earth over a period of 230 days

so as to address emerging constraints from the contaminants. Comparison between the PbS and PbSO<sub>4</sub> ( $r = 0.5$ ), PbS and PbCO<sub>3</sub> ( $r = 0.60$ ), PbS and PbCO<sub>3</sub> ( $r = 0.6$ ) treatment means showed a significant difference ( $p < 0.05$ ) while demonstrating a very close pattern of amelioration with red earth. Lead levels increased in various control samples over time after the equilibration phase with low values demonstrated in samples treated with red earth. The greatest difference of  $9.14 \mu\text{g L}^{-1}$  was observed when PbS was compared to PbCO<sub>3</sub> with the least difference of  $2.1 \mu\text{g L}^{-1}$  demonstrated when PbSO<sub>4</sub> was compared to PbCO<sub>3</sub>. The difference between the blanks (control with/without) was not significant.

#### Effect of Bone Meal Treatment on pH Values of Leachate from Samples Spiked with Lead Compounds

During the experiment a significant ( $p < 0.05$ ) effect of Red earth on pH values was observed with various lead sources (Fig. 2) indicating that Red earth was acting as an alkaline regulator. The highest pH was observed with PbCO<sub>3</sub>+RE ( $6.60 \pm 0.2$ ) while the lowest pH was observed with PbNO<sub>3</sub>-RE ( $3.40 \pm 0.15$ ). However the biggest difference indicating the effect of RE was observed with PbNO<sub>3</sub> > PbCO<sub>3</sub> > PbSO<sub>4</sub>, PbS and the control showed similar results demonstrating no significant of RE on the samples.

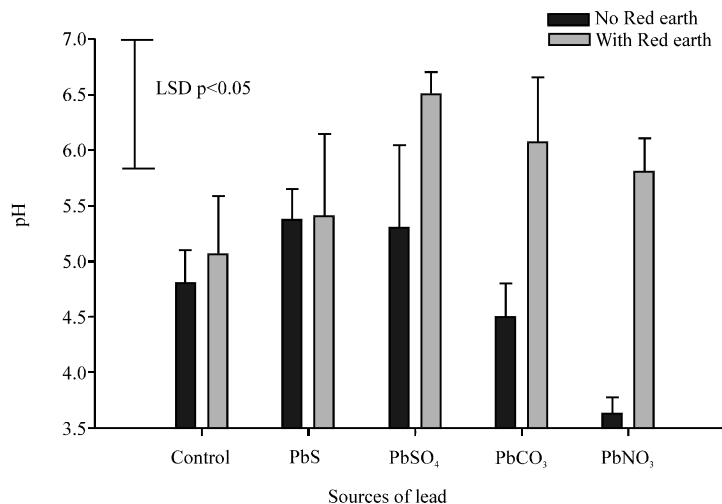


Fig. 2: Effect of Red earth treatment on pH values of various sources of lead

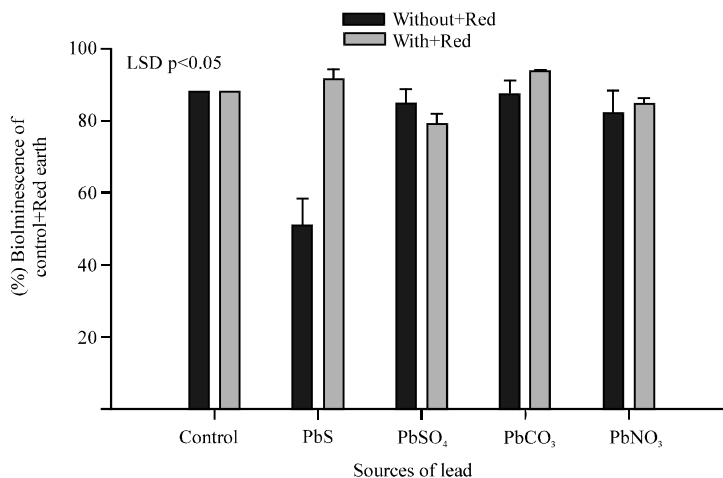


Fig. 3: Effect of red earth on Pb bioavailability in soils spiked with various lead compounds

#### Biosensor Based Toxicity Test of the Leachate

When soil samples spiked with various lead compounds (PbNO<sub>3</sub>, PbCO<sub>3</sub>, PbSO<sub>4</sub> and PbS) were treated to Red earth a significant effect was observed ( $p < 0.05$ ) (with the exception of PbCO<sub>3</sub>) (Fig. 3). The highest difference was observed with PbS where on application of red earth significantly increased percentage bioluminescence from 63%  $\pm$  7 to 100.3  $\pm$  2.9. When the differences in percentage bioluminescence between samples treated with/without Red earth were further analysed the following trend was observed; PbS > PbNO<sub>3</sub> > PbSO<sub>4</sub>.

### DISCUSSION

The alumina industry worldwide is facing a growing disposal problem of Red earth/mud creating increasingly technical, economic and environmental problems. During this study, the benefit of Red mud/earth was evaluated and its use as an ameliorant for lead contaminated soils was assessed. This



was to address contamination of soils, which is currently posing major problems all over the world. Red earth/mud (Bayer process) of bauxite minerals have the following composition (which strengthens its choice as an ameliorant during this study) according to Ciccu *et al.* (2003): SiO<sub>2</sub> = 9.58%, Al<sub>2</sub>O<sub>3</sub> = 17.91%, Fe<sub>2</sub>O<sub>3</sub> = 30.45%, CaO = 7.77%, MgO = 0.86%, Na<sub>2</sub>O = 12.06%, K<sub>2</sub>O = 0.3%, TiO<sub>2</sub> = 8.61%, P<sub>2</sub>O<sub>5</sub> = 0.2%, Cu = 35 ppm, Pb = 163 ppm, As = 62 ppm, Zn = 506 ppm, Zr = 1150 ppm, V = 1476 ppm, Ba = 206 ppm, Cd = 23 ppm, Cr = 1237 ppm, Hg = 0.7 ppm, LOI = 12.38%. pH and specific gravity were 10.67 and 2.77 g cm<sup>-3</sup>, respectively.

A preliminary assessment was made regarding the scope for Red mud usage and its possible remediation potential as an ameliorant which also explored certain associated advantages such as; non-invasive, easily available and cost effective technology. The main principles of remediation technologies in terms of contaminant treatment include isolation, immobilisation, toxicity reduction, physical separation and extraction.

Environmental pollution by some of these metals (Cu, Cd, Pb, Cr, Ni etc.) is well recognised and can be detrimental to living systems (Vinod *et al.*, 2002). Many other technique for the removal of heavy metals (e.g., mechanical *in situ* removal, ion exchange chromatography, reverse-osmosis, precipitation, adsorption etc.) in aquatic and terrestrial systems are available but are considered less efficient, costly and to some extent destructive (Bower *et al.*, 1997; Dean *et al.*, 1972). Knowledge of trace element mobility and speciation in contaminated soils is an important aspect of environmental evaluation. The addition of red earth/mud to metal contaminated soils generally causes the shifting of metals from the exchangeable to the Fe-oxide fraction and decreases acid extractability of metals (Brown *et al.*, 2005). The mechanisms proposed by Lombi *et al.* (2002) suggest that specific chemisorption and possibly metal diffusion into oxide particles could be the mechanisms responsible for the fixation of metal by red mud. When Red Earth is added to contaminated soils, they can neutralise low pH and reduce metal mobility by different physical chemical mechanisms (increase of available adsorption sites) (Lombi *et al.*, 2002).

In the current study, the reduction of Pb leached from all of the lead spiked experimental pots suggested that Red earth acted as an adsorbent when lead concentration ( $\mu\text{g L}^{-1}$ ) was compared between leachate samples spiked with lead and the controls (Fig. 1 a-e), (Fig. 2). The heterogeneous adsorbency principle in red mud is associated with its ability to bind metal ions (M<sup>2+</sup>) onto one or two types of surface sites at pH < 6.0 and less than 50% surface coverage in the form of SOM<sup>+</sup> monodentate surface complex, which results in the release of protons from the surface, effectively explaining the adsorption of metals (Apak *et al.*, 1998). This observation was supported by a report by Sauvè *et al.* (2000) who indicated that the presence of Iron oxides (Red earth/mud) and also organic matter had a high capacity to adsorb Pb and concomitantly maintain a low free Pb<sup>2+</sup> activity in solution.

The increase in pH values (>pH 5.0 of all cases, with the highest unit increase noted with PbNO<sub>3</sub>) of the leachate was demonstrated during the current study (Fig. 2) for all the treatment involving red earth. This observation related to Red earth/mud of increasing pH values was supported by Zouboulis and Kydros *et al.* (1993) who earlier reported that Red earth/mud acted simultaneously as an alkalinity regulator causing precipitation by forming insoluble hydroxides, thus acting as an adsorbent and a flocculent. The increase in pH was further reported to reduce the solubility of most Pb-bearing minerals, while increasing the adsorption affinity of iron oxides, organic matter and other adsorptive surfaces (Lombi *et al.*, 2002; Sauvè *et al.*, 2000). However increasing pH also increased Pb hydrolysis, inorganic matter ion-pair formation and organic matter solubility, potentially promoting higher dissolved concentrations of Pb (Brümmer *et al.*, 1986; Sauvè *et al.*, 1998a, b). Bruni *et al.* (2005) reported that Red earth/mud is characterised very high alkalinity and its major constituents are *crystalline hematite* (Fe<sub>2</sub>O<sub>3</sub>), *boehmite* ( $\gamma$ -AlOOH), *quartz* (SiO<sub>2</sub>), *sodalite* (Na<sub>4</sub>Al<sub>3</sub>Si<sub>3</sub>O<sub>12</sub>Cl) and gypsum (CaSO<sub>4</sub>·2H<sub>2</sub>O), with a minor presence of *calcite* (CaCO<sub>3</sub>), *whewellite* (CaC<sub>2</sub>O<sub>4</sub>·H<sub>2</sub>O) and *gibbsite* Al(OH)<sub>3</sub>.

Biosensor analysis of leachate toxicity showed that the application of red mud had rendered the leached fraction of the spiked lead less toxic, with an increase in percentage bioluminescence. The only lead compound which showed a degree of toxicity before the application of Red mud after 193 days was PbS. This could be attributed to the low solubility of lead sulphide when spiked into soil samples. Bataillard *et al.* (2003) showed that when lead was added as sulphate, between 10 and 20% of lead particles dissolved, regardless of the soil type with lead sulphide progressively oxidising over time.

Therefore red earth/mud application as an *in situ* inactivation technique (where an ameliorant was incorporated and mixed with lead contaminated soil) bound the toxic metals which essentially reduced their mobility in the soil thus reducing the contaminant leachability and bioavailability (Hartley *et al.*, 2004). Moreover, the advantages of Red mud/earth which relate to factors such as retention of nutrients on infertile sandy soils, reduction of eutrophication of rivers and water ways (Summers *et al.*, 1996a), groundwater recharge areas, improvement of pasture growth (Summers *et al.*, 1996b), plant P uptake (Snars *et al.*, 2004) and water retention in excessively drained soils (Vlahos *et al.*, 1989) outstripped the disadvantages. The disadvantages that were attributed to red earth/mud include high pH values, salinity and absence of nutrients and organic constituents that could possibly suppress revegetation (Xenidis *et al.*, 2005). However, precautionary principle is highly required as a public health concern for wide application of red earth/mud where off-site assessment is conducted before use to avoid recontamination of terrestrial or aquatic systems.

## CONCLUSIONS

The potential for red earth as a heterogeneous adsorptive, cost effective and environmental friendly remediative technique for reducing the leaching of lead from contaminated soils was established. Critical issues raised in the study showed that red mud drastically reduced lead contamination most probably due to shifting of lead from the exchangeable to the Fe-oxide fraction through specific chemisorption and diffusion mechanisms including a reduction of solubility and mobility of Pb thus reducing its toxicity as demonstrated through the bioluminescence results. Through the use of a biosensor, leachate toxicity was found to be drastically reduced, except for PbS. However, speciation pattern of different lead compounds need to be tied to bioavailability based on an appropriate time scale to reliably characterize the adsorption behaviour of lead in the presence of red earth. Areas that need to be studied and assessed (for public health concerns) in depth for wide spread application of red earth/mud (in terrestrial and aquatic ecosystems) include off-site effects of red earth, appropriate stripping methods of chemically bound contaminants (e.g., trace metals, cyanide etc).

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