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Vitellogenin Assay by Enzyme-Linked Immunosorbant Assay as a Biomarker of Endocrine Disruptor Chemicals Pollution

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Abstract: Research is ongoing to develop screening and testing programmes for endocrine disrupting effects of new chemicals and in the focus of this development are the fish test species common carp (*Cyprimus carpio*). In this study we have developed quantitative enzyme linked immunosorbent assays (ELISA) for VTS in common carp. The working range of the ELISA was 11.25 to 2,000 ng mL⁻¹ (75-25% specific binding/maximum antibody binding [B/B0]) with a 50% B/B0 intra- and interassay variation of 3.9% (n = 10) and 12.5% (n = 30), respectively. This ELISA is capable of detecting VTG as low as 6 ng mL⁻¹ and can accurately detect VTG in even 10 μL of plasma. The ELISA was applied to measurement of VTG production by male carp (*Cyprinous carpio*, Cyprinidae) fish exposure to ethynylestradiol. The results showed that the amount of VTG produced in plasma of exposed fish increased in logaritmic order comparing to the control group and the ELISA described here could be used as an indicator of water pollution to estrogenic pollutants.

Key words: Immunoassay, ELISA, endocrine disruptor chemicals, VTG

INTRODUCTION

The last two decades have witnessed growing scientific concerns and public debate over the potential adverse effects that may result from exposure to a group of chemicals that have the potential to alter the normal functioning of the endocrine system in wildlife and humans (Tsutsumi, 2000). According to WHO definition, an Endocrine Disruptor Chemical (EDC) is an exogenous substance or mixture that alters function(s) of the endocrine system and consequently causes adverse health effects in an intact organism, or its progeny, or (sub)populations (Damstra et al., 2002).

Mammalian studies have focused on the male since it is much easier to determine changes in the quality and quantity of sperm than of eggs, since eggs are few and are retained internaly (Lathers, 2002). Fish have a major advantage over mammals in studies of endocrine disruptors in that they produce large numbers of both eggs and sprm and release them to the external environment. An assessment of pollution requires a simple biomarker and VTG can be the easiest choice.

In very simple terms, vitellogenesis involves ovarian estradiol stimulating the liver to produce VTG, which in turn is incorporated into the yolk of developing oocytes. Estrogenic xenobiotics can also act on the hepatic receptors to induce synthesis of VTG (Nakari, 2004). Both male and female fish, as well as immature juveniles have hepatic estrogen receptors, but only the livers of female

fish will normally be exposed to estrogens (Versonnen and Janssen, 2004). Production of VTG by males and juveniles, or non-vitellogenic females can therefore provide a bioindicator of exposure to environmental estrogens (Hornung *et al.*, 2004).

Trout exposed to sewage effluent in English rivers had up to 100,000-fold increase in plasma VTG compared to fish upstream of the effluent (Aerni et al., 2004). A follow-up study showed VTG levels of over 20 mg mL⁻¹ 5 km downstream of the effluent (Sharpe et al., 2004). Elevated plasma VTG was also found in eels below the Paris outfall (Versonnen et al., 2004). Since eels may spend up to 20 years in the river before their spawning migration consequent reproductive disruption may account for some of the decrease in the European eel population (Versonnen et al., 2004). The flounder, as a bottom dwelling estuarine species, also provides a useful biomonitor of effluents in the rivers which may potentially affect marine fish stocks (Madsen et al., 2003). In another study, male flounder in a number of British rivers had elevated plasma VTG and, in the River Mersey, the concentrations were around 100 mg mL⁻¹ (10 million times higher than controls) (George et al., 2004).

Enzyme Linked ImmunoSorbent Assay (ELISA) using microliter plates have been described as an alternative to radioimmunoassay for the mammalian and non mammalian assays (Yamada *et al.*, 2002) and to monitor the effects of endocrine disrupting chemicals on reproductive function (Wade *et al.*, 2002) and a report (Tattersfield *et al.*, 1997)

has highlighted the need for a simple, rapid and reliable method for the measurement of hormones in this

communication we describe a general preparation procedure for the enzyme label of carp VTG which is applicable to other species. The method was applied to the determination of VTG production by ethynylestradiol, a potent xenoestrogen, in Carp (*Cyprinous carpio*) plasma.

MATERIALS AND METHODS

Chemicals and equipment: Cyprinous carpio VTG and anti-VTG antibodies were a kindly gift from: professor Oliana Carnevali, Dipartimento di Biologia MCA, University of Camerino, 62032 Camerino, Italy. The purified VTG is diluted to 10 and 50 mg mL⁻¹ with VTG storage buffer (PPBS, 0.1M, pH7.4; Potassium chloride, 27 mM; Tween-20, 0.05%; Bovine Serum Albumin, 0.2%; Sodium Azide, 5 mM; Aprotinin, 20 TIU mL⁻¹, 1% (Sigma).), divided into various volumes in eppendorf tubes and stored at -20°C. Goat F(ab')2 fragment of affinity isolated rabbit antigen specific antibody conjugated to alkaline phosphatase was obtained from Sigma Immuno Chemicals (catalogue number A-3937). obtained 4-Nitrophenylphosphate (pNPP) Boehringer Mannheim GmbH (catalogue number 107 905).

Ninety-six well polystyrene high-binding microliter plates (Costar Catalogue No. 3590) were used for the ELISA. Plates were sealed for storage with Anachem Sealplate sealing film. Plates were processed using a Chelsea Instruments shaker, an Anthos Model AW-1 plate washer and read with an Anthos Model HTII Plate Reader. 96-well low binding microliter plates (Costar Cat. No. 2501) were used for sample dilution and preparation of standard curves and AnachemÒ SEALPLATE plate sealing film, (PK/100) purchased from Anachem (UK).

Fish exposure: Four groups of six male common carp (initial mean weight: 268.9±52.5 g, mean±SD) originating from a local fish breeder farm (Mahyar company) kept in four separate tanks. They were exposed to 0, 5, 10 and 50 ng mL⁻¹ ethynylestradiol up to 30 days. Blood samples were taken at 0, 15 and 30 days of exposure and plasma kept at -20°C untill VTG assay.

Plate coating with VTG: LISA plates were coated with 50 ng of purified VTG. One hundred microliter of carbonate buffer (0.05M, pH 9.6) containing 500 ng per mL of purified standard VTG added to each well of the plate

and incubated for one hour at room temperature in a humid container and then overnight at 4°C and wash once with nanopure water. The plate was blocked by addition of a 200 µL per well of fresh VTG blocking buffer (Potassium Phosphate Buffer (PPB) 0.1M, pH7.4; Bovine Serum Albumin, 2%; Sodium Azide, 5 mM.), washed three times with wash buffer (Potassium phosphate buffered saline (PPBS), 0.02M, pH7.4; Tween-20, 0.05%) sealed with plate cover and stored at -20°C.

Standard assay procedure: Serial dilutions of purified standard VTG (2000 to 3.9 ng mL⁻¹) in 100 µL VTG assay buffer were made in the first 10 wells of rows A and B of the coated plate. The remaining 2 wells in these rows were reserved for B0 (0 pg) and Non-Specific Bound (NSB). Ten microliters of the diluted samples from sample dilution plate (2, 200 and 2000 times) was pipetted into the remaining 6 rows in duplicate. Twenty microliters of anti-VTG antisera diluted 1/15,000 (carp) in VTG assay buffer was added to all wells apart from NSB wells and made up the wells to 150 µL by VTG assay buffer. The plates sealed and incubated 2 h at room temperature in a humid chamber. Plates were then washed three times with wash buffer and 150 µL of anti-rabbit IgG Alkaline Phosphatase conjugate diluted 1/15,000 in VTG assay buffer added to each well and incubated overnight at 4°C temperature. Plate washed with wash buffer once and then twice with pNPP Substrate buffer (Diethanolamine, 0.5M, pH 9.8; MgCl, 0.5 mM), 175 mL per well of 1 mg per mL 4-Nitrophenylphosphate in pNPP Substrate buffer added and then incubated dark at room temperature until colour developed to around 1.5 ODs (normally approx. Forty five minutes but can be left up to 4 h). The reaction stopped by adding 50 mL of 3 M NaOH and read at 405 nm. Nanograms per mL were calculated for the samples from the standard curve using Stingray software (Dazdaq, Ringmer, UK).

Assay validation: A pool of male common carp (Cyprinus carpio) plasma ($500 \, \mu L$) placed in $100^{\circ}C$ water bath for 1 h to denature proteins from the plasma and then centrifuged. One hundred microliter aliquots were pipetted in duplicate into wells on two rows of a microliter plate. Pure VTG was added to one pair of wells to give a concentration of $2000 \, \text{ng/well}$, serial dilutions (x2) performed and the samples assayed according to standard procedure. The same procedure was used to assay VTG in by carp exposed to ethynylestradiol (0, 5, 10 and $50 \, \text{ng mL}^{-1}$). SPSS 10 for Windows software (SPSS Inc., 444 N. Michigan Avenue, Chicago, Illinois $60611, \, \text{USA}$)

was used for the statistical analysis (t-test) of VTG production in plasma of control fish group (0 ng mL⁻¹ ethynylestradiol) with other experimental groups (5, 10 and 50 ng mL⁻¹ ethynylestradiol) and duration of fish exposure (15 and 30 days).

RESULTS

Standard curves and assay sensitivity: Typical standard curves for VTG showed the working range of the ELISA was 11.25 to 2,000 ng mL $^{-1}$ (Fig. 1) (75-25% specific binding/maximum antibody binding [B/B0]) with a 50% B/B0 intra- and interassay variation of 3.9% (n = 10) (Fig. 2) and 12.5% (n = 30), respectively. This ELISA is capable of detecting VTG as low as 6 ng mL $^{-1}$.

Assay specificity: There was a strong positive correlation (r>0.999) between the amount of VTG added to deproteinated plasma and the amount found (Fig. 3).

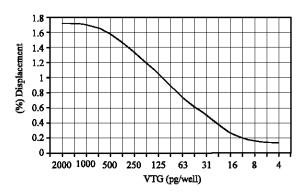


Fig. 1: Typical standard curves for ELISA of 17,20αP. Values are means of duplicate assays

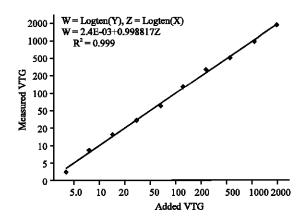


Fig. 2: Intra-assay coefficient of variance (CV) for ELISA of vitellogenin (VTG) determined from a standard curve with 10 replicates at each concentration

Application to measuring VTG production by carp exposed to ethynylestradiol: VTG was measured in duplicate by ELISA using 10 μL of the diluted plasma (2, 200 and 2000 times) from fish exposed to different concentrations of ethynylestradiol. The amounts of VTG detected by this method for fish exposed to 0, 5, 10 and 50 ng mL⁻¹ of ethynylestradio was 0, 189±25.9, 3154±684 and 24569±7698 (Mean±Standard Error of Mean for fish exposed for 15 days) and 0, 897±50.6, 12857±1235 and 86354±10423 μg mL⁻¹ (Mean±Standard Error of Mean, for fish exposed for 30 days) respectively (Fig. 4). A clear and

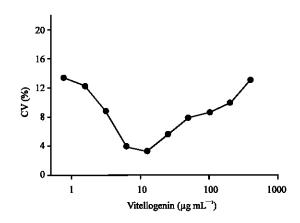


Fig. 3: Assay accuracy for vitellogenin (VTG) in deproteinized plasma. A known amount of VTG was added to an aliquot of a protein free carp plasma and 2 x serial dilutions made with it. Assays were performed in duplicate

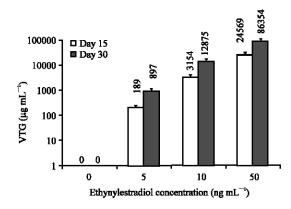


Fig. 4: The amount of vitellogenin (VTG) assayed in fish plasma exposed to different concentrations of ethynylestradiol for 30 days (vertical bars indicate SE and means significant differences in VTG induction)

potent estrogenic activity in inducing VTG by fish liver was found with increase in ethynylestradiol concentration exposure and the amount of VTG induced by different concentrations of ethynylestradiol between control (0 ng mL⁻¹) and experimental (5, 10 and 50 ng mL⁻¹) groups and between experiment groups were different significantly (p<0.05).

DISCUSSION

Examination for the effects of potential EDCs begins only at levels at which there is no apparent stress or discomfort to the animal, so that it is potentially equally able to reproduce (Ackermann et al., 2002). Reproductive dysfunction at such low concentrations can be caused either by direct action on the gametes, or indirectly by modulation of the endocrine system so that gamete development takes place during an imbalance hormonal environment (Okai et al., 2004).

In female fish, endocrine modulators can affect hypothalamic gonadotrophin releasing hormone (GnRH) secretion, pituitary gonadotrophin (GtH) release, estrogen biosynthesis and catabolism, estrogen binding to hepatic receptors, hepatic VTG production, or the feedback by steroids and VTG (Berg et al., 2004). Although hepatic production of VTG in male fish could be initiated by estrogen hormone, but no estrogenic hormones secret in male fish and no VTG production can be expected in normal life (Aerni et al., 2004). The use of in vitro fish hepatocyte cultures or in vivo fish exposure to estrogenic chemicals is shown to provide a simple and effective way to screen estrogenic activities of EDCs (Tyler et al., 1996). VTG assay in cell culture or fish plasma needs a quick, cheap, reliable and efficient method to fulfil such a duty (Fujita et al., 2004) and many researchers have already introduced different techniques to measure VTG (Anderson et al., 2003, Aerni et al., 2004), an ELISA method for assaying VTG has been described here.

Assay validation

Sensitivity: Sensitivity of the assays, which has been taken as 90% of the binding with 0 pg is just over 10 μ g for VTG. Using identical conditions for label preparation and assay we also obtained a sensitivity of 2 μ g for VTG comparable to that obtained by others (Volz and Chandler, 2004). As with radioimmunoassay, the sensitivity is a function of both the specific activity of the label and the quality of the antiserum and will therefore vary with the material available to the user. So in this study, we used

the antiserum that were available to us but have little reason to expect that any antiserum which gives good results with RIA will not give similar results to those that we describe for ELISA.

Previous reports of ELISA for VTG have used different enzyme labels (Martin-Diaz et al., 2004) and in this communication a method described previously (Cuisset et al., 1994) and the results showed it is sufficiently sensitive for the rapid measurement of VTG in common fish species such as cyprinids and present an example of such an application was described.

Accuracy and reproducibility: A major advantage of ELISA is the ability to minimise errors resulting from multiple pipetting by use of multichannel pipettes. Interand intra-assay Coefficient of Variations (CV) reported in RIA are frequently 15-20% even when measured at the steepest part of the curve. For the assays reported here we obtain intra-assay CVs around 3.9% in the steep part of the curve, rising to 15-20% at the flatter end regions (Fig. 2), while interassay variation (at 50% displacement) was around 12%, comparable to that reported by Cuisset et al. (1994). There was a highly significant correlation between the VTG added to and that measured in deproteinzed plasma (Fig. 3). A further advantage of ELISA is that sample handling can also be made more rapid and reproducible by the use of low binding microliter plates for sample storage and dispensing with multichannel pipettes.

Application to the measurement of VTG production in fish: To demonstrate an application of the ELISA method we have measured the amount of VTG produced by carp exposed to ethynylestradiol. The results confirmed previous findings that ethynylestradiol is a potent estrogenic chemical (Craft et al., 2004) (Fig. 4) and a clear and potent estrogenic activity in inducing VTG by fish was found with increase in ethynylestradiol concentration and the time of exposure (15 versus 30 days) (Fig. 4). The amounts of VTG production in fish exposed increased dramatically so the logarithmic scales used to demonstrate it in Fig 4. Such a potent induction of VTG in other studies have already been discussed (Versonnen et al., 2003).

The rapid determination of VTG from the very small volumes of blood allows the examination of water pollution with EDCs and this methodology will therefore be of great interest to workers screening the estrogenic of new chemicals and phytoestrogens.

REFERENCES

- Ackermann, G.E., J. Schwaiger, R.D. Negele and K. Fent, 2002. Effects of long-term nonylphenol exposure on gonadal development and biomarkers of estrogenicity in juvenile rainbow trout *Oncorhynchus mykiss*. Aquat Toxicol., 60: 203-221.
- Aerni, H.R., B. Kobler, B.V. Rutishauser, F.E. Wettstein, R. Fischer, W. Giger, A. Hungerbuhler, M.D. Marazuela, A. Peter, R. Schonenberger, A.C. Vogeli, M.J. Suter and R.I. Eggen, 2004. Combined biological and chemical assessment of estrogenic activities in wastewater treatment plant effluents. Anal. Bioanal. Chem., 378: 688-696.
- Anderson, M.J., D. Cacela, D. Beltman, S.J. The, M.S. Okihiro, D.E. Hinton, N. Denslow and J.T. Zelikoff, 2003. Biochemical and toxicopathic biomarkers assessed in small mouth bass recovered from a polychlorinated biphenyl-contaminated river. Biomarkers, 8: 371-393.
- Berg, A.H., L. Westerlund and P.E. Olsson, 2004. Regulation of Arctic char (*Salvelinus alpinus*) egg shell proteins and vitellogenin during reproduction and in response to 17Beta-estradiol and cortisol. Gen. Comp. Endocrinol., 135: 276-285.
- Craft, J.A., M. Brown, K. Dempsey, J. Francey, M.F. Kirby, A.P. Scott, I. Katsiadaki, C.D. Robinson, I.M. Davies, P. Bradac and C.F. Moffat, 2004. Kinetics of vitellogenin protein and mRNA induction and depuration in fish following laboratory and environmental exposure to oestrogens. Mar. Environ. Res., 58: 419-423.
- Cuisset, B., P. Pradelles, D.E. Kime, E.R. Kuehn, P. Babin, S. Davail and F. Le Menn, 1994. Enzyme immunoassay for 11-ketotestosterone using acetylcholinesterase as label: Application to the measurement of 11-ketotestosterone in plasma of Siberian sturgeon. Compar. Biochem. Physiol. Pharmacol. Toxicol. Endocrinol., 108: 229-241.
- Damstra, T., S. Barlow, A. Bergman, S. Kavlock and G. VDK, 2002. Global assessment of the state of the science of endocrine disruptors. Aquat. Toxicol., 60: 30.
- Fujita, T., H. Fukada, M. Shimizu, N. Hiramatsu and A. Hara, 2004. Quantification of serum levels of precursors to vitelline envelope proteins (choriogenins) and vitellogenin in estrogen treated masu salmon, *Oncorhynchus masou*. Gen. Comp. Endocrinol., 136: 49-57.

- George, S., M. Gubbins, A. MacIntosh, W. Reynolds, V. Sabine, A. Scott and J. Thain, 2004. A comparison of pollutant biomarker responses with transcriptional responses in European flounders (*Platicthys flesus*) subjected to estuarine pollution. Mar. Environ. Res., 58: 571-575.
- Hornung, M.W., K.M. Jensen, J.J. Korte, M.D. Kahl, E.J. Durhan, J.S. Denny, T.R. Henry and G.T. Ankley, 2004. Mechanistic basis for estrogenic effects in fathead minnow (Pimephales promelas) following exposure to the androgen 17 α-methyltestosterone: conversion of 17α-methyltestosterone to 17α-methylestradiol. Aquat. Toxicol., 66: 15-23.
- Lathers, C.M., 2002. Endocrine disruptors: A new scientific role for clinical pharmacologists? Impact on human health, wildlife and the environment. J. Clin. Pharmacol., 42: 7-23.
- Madsen, L.L., B. Korsgaard and P. Bjerregaard, 2003. Estrogenic effects in flounder Platichthys flesus orally exposed to 4-tert-octylphenol. Aquat. Toxicol., 64: 393-405.
- Martin-Diaz, M.L., S. Bamber, C. Casado-Martinez, D. Sales and T.A. DelValls, 2004. Toxicokinetics of heavy metals from a mining spill using Carcinus maenas. Mar. Environ. Res., 58: 833-837.
- Nakari, T., 2004. Estrogenicity of municipal effluents assessed *in vivo* and *in vitro*. Environ. Toxicol., 19: 207-215.
- Okai, Y., E.F. Sato, K. Higashi-Okai and M. Inoue, 2004. Enhancing effect of the endocrine disruptor paranonylphenol on the generation of reactive oxygen species in human blood neutrophils. Environ. Health Perspect, 112: 553-556.
- Sharpe, R.L., D.L. MacLatchy, S.C. Courtenay and G.J. Van Der Kraak, 2004. Effects of a model androgen (methyl testosterone) and a model anti-androgen (cyproterone acetate) on reproductive endocrine endpoints in a short-term adult mummichog (*Fundulus heteroclitus*) bioassay. Aquat. Toxicol., 67: 203-215.
- Tattersfield, L., P. Matthiessen, P. Campbell, N. Grandy, and R. Lange, 1997. SETAC-Europe/OECD/EC Expert Workshop on Endocrine Modulators and Wildlife: Assessment and Testing (EMWAT). Proceeding of SETAC-Europe/OECD/EC Expert Workshop, Veldhoven, The Netherlands, 10-13th April 1997. SETAC-Europe, Amsterdam., 1.
- Tsutsumi, O., 2000. Effects of endocrine disruptors on preimplantation embryo development. Nippon Rinsho, 58: 2464-2468.

- Tyler, C.R., T.G. Pottinger, E. Santos, J.P. Sumpter, S.A. Price, S. Brooks and J.J. Nagler, 1996. Mechanisms controlling egg size and number in the rainbow trout, *Oncorhynchus mykiss*. Biol. Reprod., 54: 8-15.
- Versonnen, B.J., K. Arijs, T. Verslycke, W. Lema and C.R. Janssen, 2003. *In vitro* and *in vivo* estrogenicity and toxicity of o-, m- and p-dichlorobenzene. Environ. Toxicol. Chem., 22: 329-335.
- Versonnen, B.J. and C.R. Janssen, 2004. Xenoestrogenic effects of ethinylestradiol in zebrafish (*Danio rerio*). Environ. Toxicol., 19: 198-206.
- Versonnen, B.J., G. Goemans, C. Belpaire and C.R. Janssen, 2004. Vitellogenin content in European eel (*Anguilla anguilla*) in Flanders, Belgium. Environ. Pollut., 128: 363-371.

- Volz, D.C. and G.T. Chandler, 2004. An enzyme-linked immunosorbent assay for lipovitellin quantification in copepods: A screening tool for endocrine toxicity. Environ. Toxicol. Chem., 23: 298-305.
- Wade, M.G., W.G. Foster, E.V. Younglai, A. McMahon, K. Leingartner, A. Yagminas, D. Blakey, M. Fournier, D. Desaulniers and C.L. Hughes, 2002. Effects of subchronic exposure to a complex mixture of persistent contaminants in male rats: Systemic, immune and reproductive effects. Toxicological Sciences: Official J. Soc. Toxicol., 67: 131-143.
- Yamada, H., R. Satoh, M. Ogoh, K. Takaji, Y. Fujimoto, T. Hakuba, H. Chiba, A. Kambegawa and M. Iwata, 2002. Circadian changes in serum concentrations of steroids in Japanese char *Salvelinus leucomaenis* at the stage of final maturation. Zoolog Sci., 19: 891-898.